Chapter 2

UREA DEGRADATION BY
ELECTROCHEMICALLY GENERATED REACTIVE CHLORINE SPECIES:
PRODUCTS AND REACTION PATHWAYS
This study investigated the transformation of urea by electrochemically-generated reactive chlorine species (RCS). Solutions of urea with chloride ions were electrolyzed using a bismuth doped TiO$_2$ (BiO$_x$/TiO$_2$) anode coupled with a stainless steel cathode at applied anodic potentials ($E_a$) of either +2.2 V or +3.0 V versus the normal hydrogen electrode. In NaCl solutions, the current efficiency of RCS generation was near 30% at both potentials. In divided cell experiments, the pseudo-first-order rate of total nitrogen decay was an order of magnitude higher at $E_a$ of +3.0 V than at +2.2 V due to the generation of dichlorine radical (Cl$_2^-$) ions. Quadrupole mass spectrometer analysis of the reactor head-space revealed that N$_2$ and CO$_2$ are the primary gaseous products of the oxidation of urea, whose urea-N was completely transformed into N$_2$ (91%) and NO$_3^-$ (9%). The higher reaction selectivity with respect to N$_2$ production can be ascribed to a low operational ratio of RCS to N. The mass-balance analysis recovered urea-C as CO$_2$ at 77%, while CO generation most likely accounts for the residual carbon. In light of these results, we propose a reaction mechanism involving chloramines and chloramides as reaction intermediates, where the initial chlorination is the rate-determining step in the overall sequence of reactions.
2.1. INTRODUCTION

Urea is the most abundant nitrogen (N) carrier in human excreta,\textsuperscript{1,3} and therefore is the dominant source of N in domestic wastewater. In spite of the relatively low toxicity of urea, there is a growing concern that excessive N loading from toilet wastewater is leading to the eutrophication of surface waters. Major cities located in sensitive coastal areas are experiencing algal blooms in estuary waters that appear to be correlated to nutrient loading from household septic systems.\textsuperscript{4} During transport through sanitary sewer systems, urea-N is readily converted to NH\textsubscript{4}\textsuperscript{+}-N via enzymatic hydrolysis by urease. In septic systems, the pH can rise to 9 due to the protonation of ammonia released during the enzymatic hydrolysis.\textsuperscript{5} Due to the heavy loading of N in human waste, various groups are exploring source separation in order to harvest the urine at a reduced volume before discharge into a large volume disposal system.\textsuperscript{2} In this respect, one can envision an on-site method for removing the N loading due to urea and thus avoiding problems resulting from the extended storage of toilet wastewater. Current demands on total maximum daily load reductions may require decentralized wastewater treatment methods, due to limited options for expanding the capacity of centralized wastewater treatment plants. The Bosch-Meiser process used to produce urea also discharges wastewater loaded with residual urea (< 30\% of products\textsuperscript{1}) and thus requires a direct urea treatment technology other than the methods of adsorption or biological treatment that are currently employed.\textsuperscript{6}

When compared to biological nitrogen removal processes (nitrification or Anammox) for urine treatment whose hydraulic retention times range up to a few days,\textsuperscript{2} electrochemical treatment can be an efficient way to accomplish urea degradation at
relatively fast reaction rates (retention time within a few hours\textsuperscript{6,7}). Operational flexibility and energy sustainability can also be achieved when using a renewable energy source (\textit{e.g.}, photovoltaic-panel\textsuperscript{8}). Botte and co-workers\textsuperscript{9-11} investigated nickel (II) hydroxide based electro-catalysts to facilitate a direct electron transfer from urea-N to anodes operating in a strong alkaline solution (\textit{e.g.}, 1 to 5 M of KOH). Other electrochemical systems have explored mixed metal oxide anodes based on RuO\(_2\), IrO\(_2\), and PtO\(_2\), to treat urea (< 0.2 M) with near equi-molar amount of chloride ion.\textsuperscript{6,7,12,13} In these systems, free reactive chlorine (HOCl or OCl\(^-\)) appeared to be the primary oxidant leading to urea degradation, while chloramines were often observed as reaction intermediates.\textsuperscript{12}

There appears to be a reasonable consensus that N\(_2\) and CO\(_2\) are the eventual products during chlorination of urea,\textsuperscript{7,13} although there are reports of N\(_2\)O (nitrous oxide) as a final N-containing product.\textsuperscript{14} However, little data can be found on the exact speciation of the potential array of gaseous products. On the other hand, we recently demonstrated that electrochemically generated chlorine radicals can rapidly reduce the chemical oxygen demand in domestic wastewater.\textsuperscript{15} In this case, the reactive chlorine species (RCS) is used as a collective term for the sum of free chlorines and chlorine radicals such as the chlorine atom (Cl\(^-\)) and chlorine radical anion (Cl\(_2\)\(^-\)).

With these considerations in mind, we now report on the details of N transformation during urea degradations via RCS generated on hetero-junction anodes functionalized with bismuth-doped titanium dioxide (BiO\(_x\)/TiO\(_2\)). Parallel analyses of products generated in the aqueous phase and gas phase allowed for mass and charge balance determinations of the nitrogenous and carbonaceous species. The effects of applied
anodic potential ($E_a$) on the rates of total nitrogen (TN) decay and products formation were also explored in order to assess the detailed roles of chlorine radicals during urea oxidation. The degradation of urea by electrochemically generated RCS seems to resemble the ‘chemical (most often using NaOCl)’ chlorination mechanisms\textsuperscript{14} (the process used to investigate the formation of chlorinated byproducts in swimming pools). However, we note that the differences in operational molar ratios of RCS to N as well as the chlorine radical generation rates change the kinetic parameters and the selectivity towards the final products in the electrolytic urea degradation.

2.2. EXPERIMENTAL SECTION

2.2.1. Preparation of BiO\textsubscript{x}/TiO\textsubscript{2} Anode and Electrolysis Cell. In a typical preparation,\textsuperscript{15,16} Ti metal sheets (0.5 mm thick, pretreated by sand-blasting, degreasing, and etching) were doubly coated sequentially with mixed metal oxide layers by repetitive thermal decomposition of Ir/Ta solution (73 mM H\textsubscript{2}IrCl\textsubscript{6} and 27 mM TaCl\textsubscript{5} in 4 M HCl; anneal at 525 °C), Sn/Bi solution (225 mM SnCl\textsubscript{4} and 12.5 mM Bi\textsubscript{2}O\textsubscript{3} in 0.5 M HCl; at 425 °C), Ti/Bi slurry (ca. 300 mM Bi-doped TiO\textsubscript{2}; at 250 °C), and Ti/Bi solution (225 mM Ti-Glycolate complex and 25 mM bismuth citrate in 75 mM NH\textsubscript{4}OH; at 425°C).

Single compartment electrolysis cells (working volume of 60 mL) were prepared using a BiO\textsubscript{x}/TiO\textsubscript{2} anode and stainless steel cathode pair (3 × 2 cm\textsuperscript{2}, 5 mm gap) with a Ag/AgCl/Sat. KCl reference electrode (BaSi Inc., USA) whose Vycor\textsuperscript{®} glass tip was 2 mm apart from the anode center. For divided cell configuration, a Pt wire encapsulated in a tube with Vycor\textsuperscript{®} glass frit separator was used as a cathode. A schematic diagram of
each electrolysis cell is shown in Figure 2.1. The frit separator reduces interferences of cathodic reactions to estimate more reliable kinetic parameters, while an increased electrical resistance of the divided cell decreases the response current density \( J \) at a given \( E_a \). All electrochemical experiments were performed under potentiostatic conditions driven by a SP-50 potentiostat (Bio-Logics, USA).

2.2.2. Potentiostatic Electrolysis of NaCl Solutions with or without Urea. The rate and current efficiency of RCS generation were estimated in the single compartment cell with 50 mM NaCl solutions. The RCS evolution rates were measured in terms of total chlorine (ClDPD) concentrations at a fixed time interval (2 min) three times. The relatively short reaction duration (6 min) was intended to minimize the oxidation or reduction of RCS, which would cause a non-linear increase of [RCS] with time.\(^{16,17}\) These experiments were repeated with \( E_a \) varying from 1.5 to 3.0 V versus the normal hydrogen electrode (NHE).
The current efficiency for the RCS generation ($\eta_{RCS}$) was estimated by the following equation:

$$\eta_{RCS} = \frac{2VF}{I} \frac{d[Cl_{DPD}]}{dt}$$

(2.1)

where, $V$ is electrolyte volume (0.06 L), $F$ is Faraday constant (96485.3 C mol$^{-1}$), $[Cl_{DPD}]$ is the concentration of RCS (M), $I$ is current (A), and $t$ is electrolysis time (sec).

The electrochemical urea degradation experiments were performed in a divided cell ($E_a$: 2.2 V or 3.0 V NHE, 3 h) or in a single compartment cell ($E_a$: 3.0 V NHE, 6 h). The reaction solution contained 41.6 mM (2.5 g L$^{-1}$) of urea in the presence of 50 mM NaCl (electrical conductivity: 5.6 ± 0.1 mS cm$^{-1}$). The initial [urea] was in accordance with a recipe for synthetic urine$^{18}$ assuming 10-fold dilution by flushing. The molar ratio of urea-N to Cl$^-$ (close to 1.7) was compatible with operational ranges in analogous literature,$^{6,12}$ while it was less than half of the observed ratios for fresh urine.$^2$ We rationalize the low urea-N/Cl$^-$ ratio by the scenario either of moderate storage before treatment (decreasing [urea]) or reuse of treated effluent as flushing water (increasing [Cl$^-$]). Urea solutions with 50 mM Na$_2$HPO$_4$ (7.4 ± 0.1 mS cm$^{-1}$) were also tested in the divided cell as control experiments, to evaluate urea oxidation by a direct electron transfer from urea to the semiconductor anode. Small aliquots of electrolyte were sampled periodically to monitor the speciation of aqueous-phase N. Gaseous products in the reactor headspace were analyzed by HPR-20 gas analysis system (Hiden Analytical, USA) with a heated quartz inert capillary inlet under continuous purging of Ar gas (ca. 7 mL min$^{-1}$).
2.2.3. **Analytical Methods.** The pH, conductivity and electrical resistance of solutions before and after the electrolysis were determined using a pH meter (Mettler Toledo, USA), a conductivity meter (VWR International, USA), and the current interruption function of the potentiostat with a current bias of 200 mA, respectively. The [urea] was determined using a hydrolysis method employing Jack Bean Urease as described previously.\textsuperscript{16} However, the measured values of [urea] were reliable only for the control experiments due to interferences of RCS with urease activity.\textsuperscript{20} The principal ions of interest (Cl\textsuperscript{−}, ClO\textsubscript{3}\textsuperscript{−}, NO\textsubscript{2}\textsuperscript{−}, NO\textsubscript{3}\textsuperscript{−}, NH\textsubscript{4}\textsuperscript{+}) were quantified by DX-500 ion chromatography system (Dionex, USA), equipped with the anion-exchange column Ionpac AS 19 and the cation-exchange column Ionpac CS 16. [TN] were determined using low range (0.5 to 25 mg N L\textsuperscript{−1}) TN reagent kits (Hach, USA) according to absorbance at 420 nm in the UV-VIS spectrophotometer (Agilent, USA) after digestion. The [Cl\textsubscript{DPD}] was measured with DPD (N,N-diethyl-p-phenylenediamine)/KI reagents (Hach, USA) based on the absorbance at 530 nm. It has been reported that, in urea chlorination, the Cl\textsubscript{DPD} accounts for chlorinated urea species as well as inorganic chloramines.\textsuperscript{21} Analogous measurements with DPD reagents (Hach, USA) were performed to quantify the fraction of free chlorine in the Cl\textsubscript{DPD}. The gaseous products underwent 70 eV electron impact ionization, and fragments (m/z 1 to 300) were analyzed using a quadrupole mass spectrometer (QMS) in a vacuum (< 5.0 × 10\textsuperscript{−6} Torr) generated by a turbo pump. The volume percent in the headspace gas was assumed to be directly proportional to the ion current measured by the QMS. Quantification of the gases of interest was based on the relative ion current intensities of characteristic m/z (14 for N\textsubscript{2}; 44 for CO\textsubscript{2}) and the volumetric flow rate of total gaseous products measured with a mass flow-meter (Bronkhorst, USA). The limit of
detection, defined as three times the standard deviation of the relative ion intensities under injection of blank gas (Ar), was estimated to be *ca.* 100 ppm for m/z 1 to 46 (except fragments from Ar) and 1 ppm for m/z 47 to 300. Additional details are given in Supporting Information.

### 2.3. RESULTS AND DISCUSSION

**2.3.1. RCS Generation on BiO<sub>x</sub>/TiO<sub>2</sub> Anode.** Figure 2.2 shows the RCS generation profiles obtained with the BiO<sub>x</sub>/TiO<sub>2</sub> anode in 50 mM NaCl solutions with circum-neutral pH. The Cl<sub>DPPD</sub> in absence of N compounds accounts for RCS, although it should exist predominantly as free chlorine due to the short life of the chlorine radicals. Cl<sup>-</sup> oxidation in terms of the specific RCS generation rate monotonically increased along with *J* as the *E<sub>a</sub>* increased from 1.5 to 3.0 V NHE (corresponding to cell voltages ranging from 2.5 to 6.0 V). The dependency of *J* on *E<sub>a</sub>* was nominally linear in dilute chloride solutions due to resistances between anode and the reference electrode; i.e., *iR*-drop attenuates the exponential response of *J* described by the Butler-Volmer formulation. The *η<sub>RCS</sub>* was below 10% at *E<sub>a</sub>* of 1.5 V NHE (slightly higher than the redox potential of the Cl<sub>2</sub>/Cl<sup>-</sup> couple), and increased to 24 – 30% at higher potentials. The losses in current efficiency (> 70%) can be dominantly ascribed to the oxygen evolution reaction. In the case of the mixed metal oxide (BiO<sub>x</sub>/TiO<sub>2</sub>) anode, the surface titanol (i.e., >TiOH) groups provide the active sites for surface hydroxyl radical formation, while the redox transitions of BiO<sub>x</sub> (III/V) or underlying IrO<sub>x</sub> (IV/VI) serve as electron sinks.
Figure 2.2. Current efficiency (circle) and rate (square) of reactive chlorine generation together with current density (triangle) on average as functions of applied anodic potential during potentiostatic electrolysis of 50 mM NaCl solutions in single compartment cell; anode: BiOx/TiO2 (6 cm²), cathode: stainless steel (6 cm²). Error bars represent the standard deviations for three samples collected with 2 min intervals.

In this study, the current efficiencies were quasi-constant at potentials higher than 1.8 V NHE. These observations have the following implications; i) the power efficiency for RCS generation should decrease along with $E_a$, and ii) one can expect a linear dependence of the RCS generation rate on $J$ under the experimental conditions of this study.

2.3.2. Aqueous and Gaseous Products Formed During Urea Degradation. Figure 2.3 shows the transformation of nitrogen during electrolysis of urea solutions with Cl⁻ in the divided cell. Urea degradation was negligible in a phosphate electrolyte or under $E_a$ of
Figure 2.3. Time profiles of (a) total nitrogen, (b) total chlorine (Cl<sub>DPD</sub>), (c) NH<sub>4</sub><sup>+</sup>, and (d) NO<sub>3</sub><sup>-</sup> concentration during potentiostatic electrolysis of 41.6 mM urea solutions with 50 mM Cl<sup>-</sup> (60 mL) in divided cell; anode: BiO<sub>x</sub>/TiO<sub>2</sub> (6 cm<sup>2</sup>), cathode: Pt wire in frit separator, applied anodic potential: 2.2 V NHE (empty circle), 3.0 V NHE (filled circle).
1.3 V NHE in a chloride background (Figure 2.4). These results indicate that a direct electron transfer from urea-N to surface sites on the anode is negligible. The [ClDPD] was consistently below the detection limit in these control experiments so that a potential interference of reactive oxygen species could be ruled out in subsequent electrolysis experiments. We recently reported\textsuperscript{15,16} that, on the BiO\textsubscript{x}/TiO\textsubscript{2} anode, the surface bound hydroxyl radicals are readily quenched by redox transitions of Bi(III/IV) to form higher oxides which prefer oxygen evolution to reactive oxygen species generation.\textsuperscript{16}

The aqueous products from the electrolytic degradation of urea included NH\textsubscript{4}\textsuperscript{+}, NO\textsubscript{3}\textsuperscript{-}, chloramines \textit{(i.e.,} NH\textsubscript{2}Cl, NHCl\textsubscript{2}, NCl\textsubscript{3})\textit{, and chlorinated urea. Concentrations of free chlorine were below the detection limit, indicating that the chloramines and chloramides primarily account for the measured ClDPD. The time profiles of the aqueous products were compatible with the previous reports on chemical urea oxidation by NaOCl.\textsuperscript{14} These
suggested that urea is sequentially chlorinated to tetrachloroureara which is oxidized further to produce CO₂ and inorganic chloramines. The reversible hydrolysis of chloramines will yield NH₄⁺ depending on the pH and the [RCS]. In addition, the inorganic chloramines are believed to undergo base catalyzed disproportionation to produce gaseous N as in breakpoint chlorination, while NHCl₂ is further oxidized to NO₃⁻. The overall chemical reaction is known to be limited by the first chlorination step of urea, whereas subsequent reactions proceed at much faster rates.

At circum-neutral pH, the nominal second-order rate constant for the reaction of urea with free chlorine was determined to be 0.63 M⁻¹ s⁻¹. In contrast, the bimolecular reaction between NHCl₂ and NCl₃ has a rate constant of 5.6 × 10³ M⁻¹ s⁻¹, while the reaction between NH₂Cl and NCl₃ has a bimolecular rate constant of 1.4 × 10² M⁻¹ s⁻¹. In this study, the rate-limiting step in the reaction sequence was found to be the initial reaction between urea and RCS as well. The evolution of gaseous nitrogen appeared to correspond directly to the ‘urea + RCS → reaction intermediates’, since the rate of urea degradation was correlated with the rate of TN decay. The sums of Cl₃DPD, NH₄⁺, and NO₃⁻ concentrations were much smaller than the reduced TN concentrations, which in turn should account for the production of gaseous N species.

Figure 2.5 shows the relative intensities of the ion fragments for the gas products in the single compartment cell experiments. Based on the NIST/EPA/NIH Mass Spectral Library (NIST11), the major peaks in the mass spectrum (m/z 1 to 46) consisted of H₂ (m/z 1 and 2), O₂ (m/z 34, 32 and 16), H₂O (m/z 16 to 19), CO₂ (m/z 44 to 46, 28, 22, 16,
Figure 2.5. Relative intensities of ion fragments in quadrupole mass spectrometer analysis of gaseous products during potentiostatic electrolysis of 41.6 mM urea solutions with 50 mM Cl⁻ (60 mL) in single compartment cell; anode: BiOₓ/TiO₂ (6 cm²), cathode: stainless steel (6 cm²), applied anodic potential: 3.0 V NHE. Error bars represent the standard deviations of periodic measurement (detection interval: 1 min).
and 12), N\(_2\) (m/z 29, 28 and 14), and Ar (m/z 40, 38, 36 and 20). From the mass spectral analysis, N\(_2\) appears to be the principal gaseous N product in urea degradation. A negligible signal at m/z 30 (i.e., N\(_2\)O) indicates that nitrous oxide is a trace-level product. On the other hand, the ratios of mean signal intensities from H\(_2\) to those from O\(_2\) were much higher than the stoichiometric ratio in water splitting (2:1). This implies that non-stoichiometric water splitting was taking place. Heterogeneous Cl\(^-\) oxidation appears to be the main scavenging reaction of oxygen generation at the electrode surface.\(^{26}\) This process in turn increases the bulk solution pH during electrolysis. Considering that the intensity of m/z 15 was below the detection limit, a stripping of dissolved NH\(_3\) is not apparent due to the low [NH\(_4^+\)] in solution (vide infra).

### 2.3.3. Urea Degradation Kinetics in a Divided Cell

As an extension of our recent work,\(^{15}\) we argue that the electrochemical transformation of urea is accelerated by generation of chlorine radical species. The mean current density at applied potential, \(E_a\), at 3.0 V NHE (83 A m\(^{-2}\)) was 3.6 times higher than at 2.2 V NHE (23 A m\(^{-2}\)). If the heterogeneous RCS generation is rate limiting, then the formation rates of NO\(_3^-\) or N\(_2\) should be proportional to \(J\), given the comparable values of \(\eta_{RCS}\) as shown in Figure 2.2. However, the apparent pseudo first-order decay rate of TN (\(k_{TN}\)) in Figure 2.3a was estimated to be 4.4 \(\times\) 10\(^{-6}\) s\(^{-1}\) and 4.0 \(\times\) 10\(^{5}\) s\(^{-1}\). The increase of \(k_{TN}\) with \(E_a\) (9 times) exceeds the proportionality as predicted by the values of \(J\) (3.6 times). In addition, the initial rates of Cl\(_{DPD}\) evolution in Figure 2.3b increased by more than an order of magnitude under the higher potential bias (5.5 \(\times\) 10\(^{-8}\) M s\(^{-1}\) and 7.8 \(\times\) 10\(^{-7}\) M s\(^{-1}\), respectively, by linear regression to the initial three data points).
Figure 2.6. Time profiles of relative intensities for ion fragments m/z 2 and 28 in quadrupole mass spectrometer analysis during potentiostatic electrolysis of 41.6 mM urea solutions with 50 mM Cl⁻ (60 mL) in divided cell; $E_a$: 2.2 V NHE (dashed line), 3.0 V NHE (solid line), anode: BiOx/TiO₂ (6 cm²), cathode: Pt wire in frit separator. Inset figure shows the relative intensities on average for ion fragments m/z 2, 28, 32, 44 for the initial 1 h.

The measurement of gaseous products in a divided cell (Figure 2.6) further supports the enhanced N₂ generation with increasing applied potential. When the $E_a$ was set at 3.0 V NHE, mass fragments appeared immediately at m/z 2 (H₂), 32 (O₂), and m/z 28 (N₂ and CO₂). When the potential was set at 2.2 V NHE, in contrast, the generation of N₂ and CO₂ were minimized, while the oxygen/hydrogen signals were found to be in good agreement with the current density. These combined observations demonstrate that the urea degradation is limited by the homogeneous reaction between urea and RCS.

The chemical reaction network for the electrochemical oxidation of urea mediated by RCS can be summarized in terms of three kinetically significant steps. They are i)
heterogeneous RCS generation, ii) the initial reactions of RCS with urea, and iii) subsequent oxidation reactions involving the initial reaction intermediate (expressed as Cl\textsubscript{DPD}) to produce N\textsubscript{2}. Given this simplification, the governing equations can be written as eq 2.2 and 2.3:

$$
-\frac{d[TN]}{dt} = 2\left(\sum k_{ii}[RCS_{i}]\right)[\text{Urea}] = k_{TN}[TN], \quad k_{TN} = \sum k_{ii}[RCS_{i}] 
$$

(2.2)

$$
\frac{d[\text{Cl}_{\text{DPD}}]}{dt} = k_{TN}[\text{Urea}] - k_{2^*}[\text{Cl}_{\text{DPD}}], \quad k_{2^*} = \sum k_{2i}[RCS_{i}] 
$$

(2.3)

where \(k_{1i}\) and \(k_{2i}\) represent the second order rate constants for the reaction of reactive chlorine species \(i\) (RCS\(_i\)) with urea and reaction intermediates (Cl\textsubscript{DPD}), respectively. The presence of the most abundant reaction intermediates should justify the usage of Cl\textsubscript{DPD} as an effective summation term. The pseudo-steady-state concentrations of RCS\(_i\) allow for the introduction of collective constants, \(k_{TN}\) and \(k_{2^*}\). Eq 2.3 corresponds to the situation in which the initial oxidation of urea is the slow rate-determining step compared with the following steps. In this case, monochlorourrea was presumed to be the primary Cl\textsubscript{DPD}, since sequential chlorination would provide more acidic amido-N moieties which undergo more facile hydrogen abstraction.\textsuperscript{14} Use of the divided cell reactor should minimize other RCS-quenching reactions, such as cathodic RCS reduction and anodic chlorate formation. For example, the final chlorate concentrations in the divided cell experiments were less than 1% of the initial chloride concentration, primarily owing to the relatively low current density (\textit{vide infra}). The above set of simplified equations describes the [TN] versus time in terms of a simple pseudo first-order decay profile
(when 2[Urea] ~ [TN]). Non-linear regressions to the observed [ClDPD] predicted $k_2^*$ of $5 \times 10^{-4}$ sec$^{-1}$ at 2.2 V NHE and $7 \times 10^{-4}$ sec$^{-1}$ at 3.0 V NHE.

The bulk solution pH increased up to 9.5 after the divided cell experiments, due to the non-stoichiometric water splitting. The pK$_a$ of HOCl (7.5) and the solution pH suggest that OCl$^-$ would be the dominant oxidant at the $E_a$ of 2.2 V NHE. However, an acidic environment in anode vicinity would allow for contributions of Cl$_2$, which was reported to have higher reactivity to urea than HOCl and OCl$^-$. Although our simplified kinetic approach could not provide rate constants of individual RCS, the collective kinetic parameters ($k_{TN}$, $k_2^*$) qualitatively indicate a shift in RCS speciation. The iR-compensated anodic potential at $E_a$ of 3.0 V NHE (Table 2.1) was greater than the equilibrium potential needed to liberate ‘free’ chlorine radical species (Cl$^-$/Cl$_2^-$). In this case, the urea degradation would be initiated by hydrogen abstraction or electrophilic addition of chlorine atom by radicals, whose bimolecular rate coefficients are expected to be much higher than those by free chlorine. Considering the [Cl$_2^-$]/[Cl$^-$] ratio with [Cl$^-$]

<table>
<thead>
<tr>
<th>Cell Type</th>
<th>$E_a$ (V, NHE)</th>
<th>$E_a - iR$ (V, NHE)</th>
<th>$J$ (A m$^{-2}$)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Divided</td>
<td>2.2</td>
<td>1.73 (3.1)</td>
<td>23.8 (20)</td>
</tr>
<tr>
<td></td>
<td>3.0</td>
<td>2.24 (4.7)</td>
<td>83.3 (14)</td>
</tr>
<tr>
<td>Single Compartment</td>
<td>3.0</td>
<td>2.05 (3.5)</td>
<td>458 (7.5)</td>
</tr>
</tbody>
</table>

Table 2.1. The ohmic drop compensated anodic potential ($E_a - iR$) and current density ($J$) on average during the electrolysis of 41.6 mM urea solutions with 50 mM chloride (60 mM) under variable cell types and applied potentials ($E_a$).
of 50 mM ($7 \times 10^3$ using $K_{eq}$ of $1.4 \times 10^3$ M$^{-1}$), the dichlorine radical anion is expected to be the major radical species under our operating conditions.$^{26}$ The reactivity between urea and Cl$_2 \cdot$ has not been reported in literature. As a reference, apparent rate constants of free chlorine reacting with primary amines are in the order of $10^4$ M$^{-1}$ s$^{-1},$ while the reactivity of dichlorine radical has been reported up to $10^7$ M$^{-1}$ s$^{-1}.^{27}$ Nevertheless, the increases of $k_{TN}$ and $k_2^\ast$ with $E_a$ were within similar orders of magnitude, indicating that the roles of free chlorine should not be ignored even under the radical generation. The Cl$_2 \cdot$ radicals in the aqueous phase are expected to be rapidly disproportionated to free chlorine with diffusion-limited second-order rate constants ($> 10^9$ M$^{-1}$ s$^{-1}).^{26}$

2.3.4. Urea Transformation Characteristics in Single Compartment Cell. Additional experiments were performed in a single compartment cell with a paired electrode module ($E_a$: 3.0 V NHE) to investigate the transformation efficiency in a practical setup. As shown in Figure 2.7, the [ClDPD] followed a typical profile of the breakpoint chlorination; i.e., an initial rise with chloramines/chloramides production, subsequent decrease by chloramine disproportionation to N$_2$ until the breakpoint is achieved after 4 h, followed by subsequent further increase of free chlorines. Free chlorine concentration was inappreciable until 4 h of electrolysis, while almost superimposable to [ClDPD] for the last sample (after 6 h). [NH$_4^+$] showed a peak (0.7 mM) at 3 h of electrolysis and decayed to be negligible along with the dismutation of chloramines. Consequently, the initial urea-N was almost completely transformed into nitrate and N$_2$ after 6 h, where the residual [TN] ($\sim 8$ mM) roughly coincided with the [NO$_3^-$]. Figure 2.7c shows molar flow rate of N$_2$ and CO$_2$ quantified by volumetric gas flow rate and the observed QMS output signals of
Figure 2.7. Time profiles of total nitrogen (circle), total chlorine (triangle), ammonium (square), and nitrate (diamond) concentration together with molar flow rate of N$_2$ (solid line) and CO$_2$ (dotted line) during potentiostatic electrolysis of 41.6 mM urea with 50 mM Cl$^-$ (60 mL) in single compartment cell; anode: BiO$_x$/TiO$_2$ (6 cm$^2$), cathode: stainless steel (6 cm$^2$), applied anodic potential: 3.0 V.
characteristic m/z ratios. The generation rates of these major gas products followed exponential relationships as predicted by the pseudo first-order kinetic analysis. A mass balance analysis for N indicated that 5 mmol of urea N is transformed into 4.56 mmol N as N\textsubscript{2} (91\%) and 0.45 mmol N as NO\textsubscript{3}\textsuperscript{-} (9\%). The total amount of CO\textsubscript{2} in the gas products was quantified to be 1.92 mmol, somewhat smaller than the carbon input (2.50 mmol). Furthermore, the signal response at m/z 28 was greater than the expected sum by the characteristic m/z signals of N\textsubscript{2} and CO\textsubscript{2}. These observations indicate that CO is being formed as a minor carbon-containing product.

Greater selectivity for gaseous N species than for NO\textsubscript{3}\textsuperscript{-} has been reported previously in the electrochemical oxidation of urea\textsuperscript{5,7,8} and NH\textsubscript{4}\textsuperscript{+}\textsuperscript{23,28}. A lower selectivity toward NO\textsubscript{3}\textsuperscript{-} is beneficial with respect to the chlorine demands exerted by urea which, in turn, should be proportional to the energy consumption in electrochemical systems. Stoichiometric chlorine demands range from 3 to 8 mol mol\textsuperscript{-1} of Cl\textsubscript{2}/urea depending on the selectivity. On the contrary, chemical urea chlorinations\textsuperscript{3,14} under initial chlorine dosage exceeding 1 mol mol\textsuperscript{-1} of Cl\textsubscript{2}/urea have reported much higher relative yields of NO\textsubscript{3}\textsuperscript{-} (up to 40\%). The nitrate yields appeared to monotonically increase with the applied chlorine dosage (3 ~ 10 mol mol\textsuperscript{-1} of Cl\textsubscript{2}/urea), with observed chlorine demands up to 5 mol mol\textsuperscript{-1} of Cl\textsubscript{2}/urea.\textsuperscript{3} The NO\textsubscript{3}\textsuperscript{-} is believed to be produced from dichloramine oxidation by RCS\textsubscript{12,22}, which should be more sensitive to the RCS concentration than N\textsubscript{2} generation reactions among chloramines. Therefore, a low yield for NO\textsubscript{3}\textsuperscript{-} in the electrolytic urea degradation can be attributed to a presumable low RCS/urea ratio with pseudo steady-state concentration of RCS. The present results have some practical implications for control of selectivity towards N\textsubscript{2} and NO\textsubscript{3}\textsuperscript{-} in treatment of N-containing wastewater. The
electrochemical urea chlorination with greater selectivity to N₂ can be broadly deployed for remediation of N-rich wastewater (black water from toilet and livestock industry, landfill leachate), aiming at N loading reduction. When focusing on N recovery for fertilizer production, increasing $J$ is expected to elevate the nitrate yield, which is consistent with our observations in this study. In the divided cell experiments under lower $J$ (Figure 2.3), the nitrate yield was estimated to be 3.7% at $E_a$ of 3.0 V NHE. An external RCS generation in a multi-component reactor with flow regulation can be another feasible approach to enhance the nitrate yield.

Despite the higher TN decay rate with respect to the electrolysis time, the magnitude of the TN decrease per unit of passed charge was lower in the single compartment cell than in the divided cell (Figure 2.8) at a given $E_a$. The lower current efficiency in the absence of the fritted-glass separator can be attributed to quenching reactions of RCS such as cathodic reversion to Cl⁻ and further oxidation to ClO₅⁻. When the total passed charge (5936 C) in the single compartment cell experiment was compared with the theoretically required charge for the observed products, the current efficiency was estimated to be 22% for N₂ (6 e⁻ transfer per molecule), 5.9% for NO₃⁻ (8 e⁻), 7.0% for ClO₃⁻ (6 e⁻), and 1.4% for residual ClDPD (2 e⁻), respectively. The sum of anodic current efficiency (36%) slightly exceeds the $\eta_{RCS}$ measured in NaCl solutions (30%, Figure 2.2), supporting the hypothetic CO generation as an electron source. The nominal energy consumption in the single compartment cell electrolysis (6 h) was 186 kJ L⁻¹, considerably greater than in biological nitrogen removal process (ca. 5 kJ L⁻¹). The relatively intensive energy requirement of electrochemical approaches can be mitigated
by their compatibility with renewable energy sources. For example, our recent report\textsuperscript{15} estimated that about 3.5 m\textsuperscript{2} of a photovoltaic panel can achieve break-even energy balance in an electrochemical system designed to treat wastewater from a community toilet for 30 users.

2.3.5. Byproducts Formation. The electrolytic urea degradation in the single compartment cell showed that ClO\textsubscript{3}\textsuperscript{−} is a dominant aqueous-phase byproduct, most likely formed by the heterogeneous oxidation of free chlorine (HOCl and ClO\textsuperscript{−}).\textsuperscript{15,16} The [ClO\textsubscript{3}\textsuperscript{−}] increased almost linearly with the electrolysis time (or specific amount of charge passed) until the breakpoint, after which the generation was accelerated (Figure 2.9). The [ClO\textsubscript{3}\textsuperscript{−}] profiles versus specific passed charges did not show a noticeable deviation between the
Figure 2.9. Evolutions of ClO$_3^-$ concentration as functions of electrolysis times and (inset) specific passed charges during potentiostatic electrolysis of 41.6 mM urea solutions with 50 mM Cl$^-$ (60 mL) in a divided (triangle) and single compartment (square) cell; $E_a$: 3.0 V NHE, anode: BiO$_x$/TiO$_2$ (6 cm$^2$).

single component and divided cell. Therefore, the current efficiency of chlorate production is likely more sensitive to the relative amount of potential electron donors to RCS, than it is to $J$. In this study, the ClO$_3^-$ was a stable byproduct without an appreciable concentration of perchlorate ion (ClO$_4^-$). The ClO$_4^-$ has often been observed by oxidation of ClO$_3^-$ on a boron doped diamond anode.$^{29}$ In light of the standard redox potential of the ClO$_4^-$/ClO$_3^-$ couple (1.19 at pH 0), the production of ClO$_4^-$ is likely to be kinetically limited in our operational condition.

This study could not clearly identify the volatile chlorinated compounds other than NH$_2$Cl from a time lapse QMS measurement (detection interval: 1 min) of the gas products (Figure 2.5b). The molecular ion signals from m/z 72 to 300 were always lower than the limit of detection (1 ppm). Weak peaks at m/z 51 and 53 could be assigned to
NH$_2^{35}\text{Cl}$ and NH$_2^{37}\text{Cl}$. Their maximum ion intensities were always observed at the start-up and then decayed to below the detection limit within 1 h. The peak cluster at m/z 48 to 52 and 64 corresponded to chloromethane (CH$_3^{35}\text{Cl}$ and CH$_3^{37}\text{Cl}$), chloroethane (C$_2$H$_5^{35}\text{Cl}$), and their fragments. The other detected signals could not be explained in terms of the chlorinated compounds in the database, although they most likely followed fragmentation characteristics of condensed alkanes and alkenes (C$_4$H$_8$, C$_4$H$_{10}$, C$_5$H$_{10}$, and C$_5$H$_{12}$), methyl isocyanate (CH$_3\text{NCO}$), oxalic acid (H$_2$C$_2$O$_4$), and acetic acid (CH$_3$COOH). However, these compounds with reduced oxidation states of carbon are not likely to evolve from urea chlorination. Despite a lack of mass spectral data in the library, we suspect that these peak clusters with minor fractions might be ascribed to chlorinated carbamic acids or chlorinated hydrazines as potential reaction intermediates (vide infra).

Volatile disinfection byproducts reported in chlorinated swimming pools$^{30-32}$ include inorganic chloramines, C(NCl, CHCl$_3$, nitriles, and halonitroalkanes, depending on precursor composition and applied Cl$_2$/N ratio.$^{32}$ Owing to the simple molecular structure of urea with fully oxidized carbon, volatile byproducts from chemical urea chlorination were almost exclusively chloramines.$^{30,33}$ Although the appreciable chlorinated byproducts from RCS mediated urea oxidation were limited to ClO$_3^-$ and NH$_2$Cl in this study, a wide array of organics present in human wastes pose a risk of toxicity from other chlorinated byproducts. The harmful byproducts potentially formed during electrochemical treatment of human wastes may limit the reusability of the effluent to non-potable purposes and require post-treatment units. Further investigation is necessary to identify the matrix of by-products and their toxicity in more realistic situations.
Figure 2.10. Proposed reaction pathways of urea degradation by electrochemically generated reactive chlorine species (RCS); (a) Cl\(^-\) oxidation on BiO\(_x\)/TiO\(_2\) anode to produce RCS and ClO\(_3\)^-, (b1) chlorinated carbamic acids mediated chloramines and CO\(_2\) (CO) generation, (b2) chlorinated hydrazine mediated chloramines and CO\(_2\) generation, and (c) chloramine mediated N\(_2\) and NO\(_3\)^- generation. Bold arrows in (b) and (c) represent redox reactions involving the RCS while the final products are specified in bold characters.

**2.3.6. Urea Degradation Pathways.** Based on the experimental findings, we propose reaction pathways between urea and electrochemically generated RCS as shown in Figure 2.10. The evolution of [Cl\(_{DPD}\)] supports chloramines and chloramides as the reaction intermediates. The initial chlorination step appears to be the rate determining since the electron-withdrawing effects of substituted Cl would reduce the bond strength of neighboring amido hydrogen. The overall rate appears to be accelerated by chlorine radicals via hydrogen atom abstraction and rapid addition. A radical driven cleavage of carbonyl double bond or resonance stabilized C-N bond should be minimized in the
initial step, since the dominant carbon end product was observed to be CO$_2$. The dichloroureia or even monochloroureia will undergo rapid hydrolysis rather than further chlorine substitution,$^{34}$ via a pattern similar to base catalyzed haloform reaction, to produce mono-/di-chloramine and carbamic acid. Further chlorinations of the carboxylic intermediates produce more chloramines, CO$_2$ and presumably CO as a minor product. A generation of hydrazine by a Hoffmann rearrangement of monochloroureia is also feasible,$^6$ where CO is unlikely to form. The inorganic chloramines undergo a breakpoint chlorination pathway to produce dominantly N$_2$ or further oxidation to NO$_3^-$ with a lower yield.

2.4. SUPPORTING INFORMATION

2.4.1. Quantification of Molar Flow Rates of Gas Products. The m/z values of 14 and 44 were chosen as characteristic m/z for N$_2$ and CO$_2$, respectively. Potential interferences from CO$_2$, CO and N$_2$O prohibited the ion intensity of m/z 14 to be used for N$_2$ quantification. Based on the negligible signal at m/z 30, an interference of N$_2$O to the signal at m/z 44 was assumed to be marginal. The signal fraction of each characteristic m/z was calculated by the ratio to the sum of ion intensities (m/z 1 to 300). Based on the fragmentation profiles of standard N$_2$ and CO$_2$ provided by the supplier, the volume fractions were calculated by the following equations:

$$F_{N_2} = \frac{1080}{72} \frac{I_{14}}{\sum I_i} F_i \quad (m/z \ 28:14:29 = 1000:72:8)$$ (2.4)
\[ F_{CO_2} = \frac{1199}{1000 \sum_{i} I_i} \cdot F_T \quad \text{(m/z 44:28:16 = 1000:114:85)} \quad (2.5) \]

where, \( I_i \) is ion intensity of fragment m/z \( i \), while \( F_{N2}, F_{CO2} \), and \( F_T \) represent the molar flow rate of \( N_2, CO_2 \), and total gaseous products, respectively. \( F_T \) was calculated by converting the observed volumetric flow rate using ideal gas law, while considering conversion factors of the major gas components (1.4 for Ar, ca. 1 for H\(_2\), O\(_2\), N\(_2\), H\(_2\)O).

\[ F_T = \left( \frac{X_{Ar}}{1.4} + \frac{1 - X_{Ar}}{1} \right)^{-1} \frac{P}{RT} Q_{obs} \quad (2.6) \]

where, \( X_{Ar} \) is the volumetric fraction of Ar, \( Q_{obs} \) is the observed volumetric flow rate of total gaseous products, \( P \) is the operational pressure (1 atm), \( R \) is the gas constant, and \( T \) is the operational temperature (293 K).

**2.4.2. Measurement of Urea Concentration.** Figure 2.11 shows the evolution of the urea concentration under variable current density conditions, which was measured by urease hydrolysis method.\(^{16}\) The measured urea concentration dropped off for the first sample under potential bias, whose further decrease was marginal. The required charge for the decrease in urea concentration for the initial 10 min exceeded the total passed charge, even when assuming one electron transfer for the urea transformation. The presence of reactive chlorine species (RCS), even with a trace concentration, would inhibit the activity of the enzyme urease.\(^{20}\) In addition, the RCS was found to interfere in the spectrometric urea analysis by diacetylmonoxime.\(^{35}\) The color generation by the reagent addition appeared to increase with the electrolysis time.
Figure 2.11. Time profiles of urea concentration during potentiostatic electrolysis of 41.6 mM urea solutions with 50 mM Cl\(^-\) (60 mL) as control; \(E_a\): (circle) 2.2 V NHE in divided cell, (triangle) 3.0 V NHE in divided cell, (square) 3.0 V NHE in single compartment cell.

2.5. REFERENCES


(34) Chattaway, F. D. The action of chlorine upon urea whereby a dichloro urea is produced. *Proceedings of the Royal Society of London Series a-Containing Papers of a Mathematical and Physical Character* **1908**, *81* (549), 381-388.